



# Habitat use and selection patterns inform habitat conservation priorities of an endangered large carnivore in southern Europe

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ABSTRACT: Understanding the habitat use and selection patterns of endangered species is essential in developing management measures that will protect critical habitat and mitigate human-wildlife conflicts. This understanding is particularly important in areas with high anthropogenic pressures. To understand the ecological role of various habitat types in the conservation of an endangered large carnivore in southern Europe, with its distinct environmental conditions and predominantly anthropogenic landscapes, we studied 18 GPS-collared brown bears Ursus arctos in Greece. We examined the use and selection of habitats according to age and sex categories and behavioral status during 5 ecologically defined seasons. Areas with rough terrain were identified as important refuge areas and were used by all bears in late hyperphagia and emergence. All bears used areas closer to human-related habitat features during the night. Habitat selection was positive for areas with rough terrain and naturalized (i.e. abandoned or not intensive) crops and areas close to water courses, while high-altitude areas and roads were avoided. The selection or avoidance of other habitats varied across bear categories and between stationary and moving behavior. We recommend that the results of the study be used to develop guidelines for species conservation and allow for prioritizing management actions that will promote the conservation of bears in Greece. In particular, the habitat use patterns provide information on how to limit interactions between humans and bears in space and/or time, while the habitat selection patterns indicate suitable habitats that should be protected/improved based on their importance and ecological role for the species.

KEY WORDS: Brown bear · Ursus arctos · Greece · Habitat selection · Habitat use · Species recovery

#### 1. INTRODUCTION

The thorough understanding of animal habitat use and selection patterns is essential in developing effective management measures to protect critical habitats, safeguard the necessary resources for a species' persistence, mitigate its potential negative interactions with humans and ultimately contribute to the conservation of endangered species (Morrison et al. 2006). This is particularly important for large carnivores, especially in anthropogenic areas, where habitat alteration may force animals to utilize subop-

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timal habitat types (Martínez-Abraín & Jiménez 2016). The study of how animals use habitats through time and across space (i.e. habitat use) and the types of habitats that are likely to be used in relation to their availability (i.e. habitat selection; Garshelis 2000), provides information on critical habitats that can be used to prioritize conservation efforts (Schofield et al. 2007) in the European context of anthropogenic landscapes (Linnell et al. 2008).

Habitat use and selection processes are inextricably linked to the characteristics of animal movement (Martin et al. 2009), behavior (Moe et al. 2007) and the way that an animal perceives the habitat or evaluates the availability of resources (Manning et al. 2004). Therefore, dynamic approaches to define habitat sampling units (i.e. resource units; RUs), such as dynamic Brownian bridge movement models (dBBMMs) (Kranstauber et al. 2012) are currently among the state-of-the-art methods to evaluate the factors influencing wildlife habitat ecology (Byrne et al. 2014, Hinton et al. 2015, 2016). According to this approach, resource units can be defined as the areas where animals are or could be present during a certain time frame based on their movement patterns; the intensity of use of these RUs, also known as the utilization distribution (UD; Worton 1989), reflects how animals select for or against a certain habitat feature when encountering it (Dugatkin & Reeve 2000).

Brown bears Ursus arctos are globally considered by the IUCN as species of Least Concern. They are the only ursid in Europe, where several populations, especially those inhabiting the southern parts of the continent, are small, isolated and threatened by habitat loss and fragmentation and by human-bear conflicts (Penteriani et al. 2020, Swenson et al. 2020). Brown bears in Greece present an interesting case for highlighting how the deeper understanding of the habitat ecology of a large carnivore can help inform conservation priorities of an endangered species. In Greece, numbering fewer than 500 individuals (Karamanlidis et al. 2015), and considered to be endangered (Mertzanis et al. 2009), brown bears reach their southernmost European distribution. Their range in Greece has been expanding in recent years (Bonnet Lebrun et al. 2020), and their population has been recovering, both demographically (Karamanlidis et al. 2015) and genetically (Karamanlidis et al. 2018). This has been partly facilitated by specific behavioral adaptations in the overall movement patterns of the species (de Gabriel Hernando et al. 2020) combined with the rewilding of the countryside after the rural abandonment by humans (Mertzanis et al. 2020).

Brown bear habitat use and selection in Europe has been systematically studied, mainly in the northern parts of the continent (Swenson et al. 2020). Throughout northern and central Europe, human activity (Martin et al. 2010) and environmental characteristics (e.g. terrain roughness) have been shown to have varying effects on habitat use and selection, depending on bear age and sex (Nellemann et al. 2007), time of day (Moe et al. 2007), season (Pop et al. 2018) and behavioral status (i.e. stationary vs. moving behaviors; Moe et al. 2007). In contrast, our understanding of the habitat ecology of brown bears in southern Europe (i.e. Mediterranean region), with its distinct environmental conditions and high anthropogenic pressures, is still incomplete (but see e.g. Mertzanis et al. 2008, Piédallu et al. 2019). In addition, rear-edge populations (i.e. animal populations residing at the lowlatitude distribution limits), such as the brown bear population in Greece, are of paramount importance for the conservation of a species (Hampe & Petit 2005).

To understand the habitat ecology of brown bears in Greece, we examined the circannual and circadian habitat use and habitat selection patterns of individual GPS-collared brown bears across space and time. Our main objective was to gain insights into the ecological role of various habitat types for the species, thus providing temporal and spatial information that can be used to identify conservation priorities for an endangered brown bear population living in an anthropogenic landscape.

# 2. MATERIALS AND METHODS

## 2.1. Study area and animals

The study was carried out in the western part of the brown bear range in Greece, an area that extends approximately over the 250 km long Greek part of the Pindos mountain range (40°00′ N, 21°03′ E) (Fig. 1). The study area forms a mosaic of elevations and typical Mediterranean habitats and is characterized by long, arid summers and short, but cold and rainy winters. Approximately 19% of the study area is covered by mature broadleaf forest, dominated by oak (Quercus sp.) and beech (Fagus sp.), 7% is mature pine (*Pinus* sp.) and fir (*Abies* sp.) coniferous forest, and 5 % is mixed forest. Transitional woodland and shrublands with junipers (Juniperus sp.), European box Buxus sempervirens and other sclerophyllous species occupy approximately 20%, natural grasslands and pastures occupy approximately 15%, and rocky outcrops and bare grounds approximately

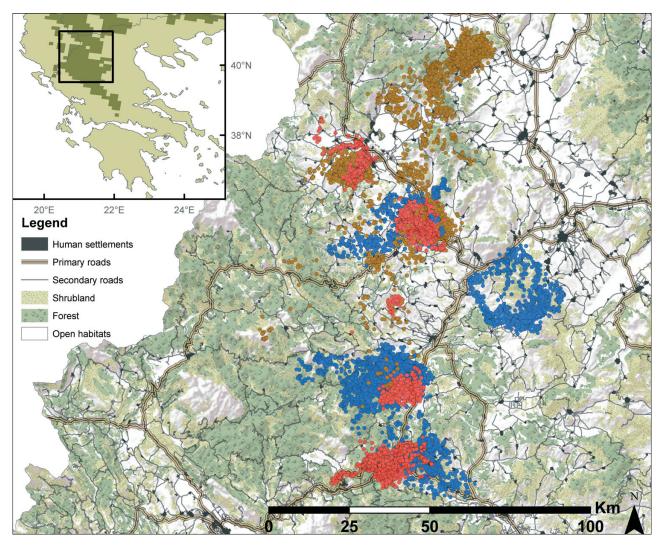


Fig. 1. Main habitat characteristics of the study area, indicating the GPS locations of collared adult female (red dots), adult male (blue dots) and subadult (brown dots) bears used in this study. The inset presents the approximate brown bear distribution in Greece, according to Chapron et al. (2014)

2% of the habitat. Human-occupied landscapes account for 32% of the habitat. Near rural areas, natural habitats are disrupted by naturalized crops (12%), which are mosaics of agricultural patches where low-intensity or abandoned crops mixed with fruit trees, shrubs and open woodlands. The main floodplains and the surroundings of human settlements are covered by intensive wheat and corn crops (20%). Elevations range from 194 m above sea level (a.s.l.) up to the peak of Mount Smolikas at 2637 m a.s.l. Compared to other bear ranges in Europe, the intensity of human activity in the study area, as measured by population density, is, on average, relatively high (i.e. 79 people km<sup>-2</sup>).

We used Aldrich leg-hold snares (Johnson & Pelton 1980) to catch bears and monitored the traps using very

high frequency alarms that were checked every 2 h and visually checked every morning. Bears were tranquilized with an initial intramuscular injection (3 ml) of a mixture of tiletamine and zolazepam (Zoletil 50; Virbac) using a CO<sub>2</sub> injection rifle (Model I.M.; DANIN-JECT) and a second injection by hand of 2 ml Zoletil 50.

We considered bears 2–3 yr of age to be subadults and older bears to be adults and assigned all individuals to 3 categories (i.e. adult females, adult males and subadult males). We fitted 3-D GPS collars (Simplex, Televilt; GPS Plus, Vectronic Aerospace) to 6 adult females, 6 adult males and 5 subadult males (2003–2013) using multilayered cotton spacers destined to break and release the collar in 1–2 yr. In addition, we collared 1 subadult female that accompanied her mother, who was not collared during the

entire study period; this was the only female-dependent cub pair in the study. Although habitat use and selection of females with dependent cubs might differ from those of lone females due to their smaller home ranges (Dahle & Swenson 2003), we included this female-cub pair within the adult female category because their movement patterns were more similar to this group than to adult or subadult males (de Gabriel Hernando et al. 2020). The GPS fix rate varied from 1 to 3 h (8-24 fixes d<sup>-1</sup>). Seven bears (4 adult females, 2 adult males and 1 subadult male) were also fitted with a dual-axis motion sensor (Ziółkowska et al. 2016). We collected a total of 37 499 GPS fixes over 3072 tracking days for 7 adult females, 6 adult males, and 5 subadult males (for details see Table S1 in the Supplement at www.intres.com/articles/suppl/n044p203\_supp.pdf.

#### 2.2. Data treatment

To define the resource units for the habitat use and selection analyses, we fitted a dBBMM to the complete movement path of each animal by calculating the 50 % UD contours and the Brownian motion variance  $\sigma_{\rm m}^2$  for all the individual time steps within each

movement path (packages 'move' [Kranstauber et al. 2019] and 'moveud' [Collier 2013] in the R statistical computing environment [R Core Team 2013]; see Text S1 for further details). Following the procedure described by Byrne et al. (2014), we then indexed each UD contour with the time of the first location belonging to that individual time step. Time steps were classified according to behavioral status as stationary (i.e. sleeping, resting, feeding, etc.) or moving (i.e. walking, running or foraging), based on their Brownian motion variance (de Gabriel Hernando et al. 2020). Finally, we selected 12 habitat variables (Table 1) that were converted to raster maps and intersected with the UD time step contours to calculate average values.

Distance-based habitat variables (Table 1) were calculated by creating path distance raster layers using a digital elevation model at 30 m resolution (USGS 2006) as a vertical factor. Path distance to polygon features was calculated from the feature edge towards the exterior (i.e. negative proximity values) and the interior (i.e. positive proximity values) of the polygon. Path distances to polyline layers (i.e. roads and water courses) were converted to negative proximity values. Roughness was calculated as the difference in elevation between neighboring

Table 1. Details of the habitat variables used in the present study

Variable	Description	Source
Altitude Roughness	Elevation above sea level calculated from digital elevation model at 30 m resolution  Terrain roughness index (i.e. difference in elevation between neighboring cells; Riley et al. 1999), calculated from a digital elevation model at 30 m resolution	Shuttle Radar Topography Mission (SRTM) digital elevation model (USGS 2006)
Forest (proximity) Shrubland (proximity) Grassland (proximity) Naturalized crops (proximity) Intensive crops (proximity)	Path distance to forested areas, including broad- leaved, coniferous or mix forests Path distance to transitional shrubland or sclerophyllous vegetation Path distance to grassland and pasture Path distance to naturalized crops or low- intensity agricultural patches Path distance to intensive crops	Modified from Corine Land Cover 2006 (CLC) seamless vector data, Version 16 (European Commission, 04/2012) (https://land.copernicus.eu/pan-european/corineland-cover/clc-2006)
Water courses (proximity) Human settlements (proximity) Primary roads (proximity) Secondary roads (proximity) Unpaved roads (proximity)	Path distance to permanent water courses (excluding lakes) Path distance to human settlements (villages and towns) Path distance to motorways or primary roads often linking larger towns Path distance to secondary or tertiary road often linking smaller towns and villages Path distance to minor roads, generally unpaved	Modified from OpenStreetMap Data in layered GIS format (www.openstreetmap.org/#map=6/38.359/23.810)

cells (Riley et al. 1999). All variable calculations were performed with the Spatial Analyst toolbox in ArcGIS 10 (Environmental Systems Research Institute). The absolute Pearson correlation coefficient for any pair of variables was less than 0.50, with the exception of the pairs 'altitude–proximity to intensive crops' (R = -0.58) and 'proximity to settlements–proximity to secondary roads' (R = 0.67). The variance inflation factor was below 4.0 for all variables, suggesting low collinearity effects (Heiberger & Holland 2004).

To analyze habitat use and selection across seasons, we used the ecologically defined seasons for bears in Greece identified by de Gabriel Hernando et al. (2020), i.e. 'emergence' (EM; 1 March–21 April), 'mating' (MA; 22 April–21 June), 'post-mating' (PM; 22 June–7 August), 'early hyperphagia' (eHY; 8 August–7 October) and 'late hyperphagia' (lHY; 8 October–15 December). We did not include data from the denning period (16 December–28 February) in our analyses due to a small sample size of successful GPS fixes.

#### 2.3. Data analysis

To study the circannual and circadian habitat use patterns of brown bears in Greece, we considered the 50% dBBMM time step UD contours of each individual as the used resource units. For each bear category and each of the 12 habitat variables (i.e. response variables), we fitted a generalized additive mixed model (GAMM) to test the effects of the predictor variables (i.e. day of the year for circannual and hour of the day for circadian) while applying a cyclic cubic spline smoother to them. For the circannual analyses, we included the individual bear ID-year (i.e. bears monitored for more than 1 yr were considered different sample units to account for annual variations within individuals) as a random effect, while for the circadian analyses, we included season as a random effect.

To study habitat selection, we followed the Design III approach, i.e. the animals were identified and both the use and the availability of habitat were measured (Manly et al. 2002). Available RUs for each bear were obtained by calculating the 50% dBBMM time step UD contours of the alternative random trajectories that were created within the annual 100% minimum convex polygon of each individual (R package 'adehabitatLT', Calenge 2006). We started the random trajectories at the capture location (Calenge et al. 2009) or at the first

location during emergence if a bear was monitored for more than 1 yr. We then followed the same pattern of distances and times between successive locations as the actual bear trajectories according to the season, but randomized the turning angles. For each bear category and behavioral status, we fitted a separate generalized linear mixed model (GLMM) with a binary response (0 = available RUs; 1 = usedRUs) and a logistic link using 250 randomly selected and stratified by season used and available RU records for each bear. We tested all additive combinations of the 12 habitat variables (R package 'MuMIn', Bartón 2019) including the bear ID and season as crossed random effects. For each bear category and behavioral status, we selected the model with the lowest value of Akaike's information criterion corrected for small sample sizes (AICc) and all of its significant variables (p  $\leq$  0.05) as the best-fitting model supporting habitat selection while considering the standardized coefficient estimates as indicators of habitat selection (i.e. positive estimates) or avoidance (i.e. negative estimates). The relative order of importance for each variable was calculated based on the  $\Delta AIC_c$  between the full model and each model for which the target variable was removed.

#### 3. RESULTS

## 3.1. Circannual and circadian habitat use

Circannual habitat use patterns differed across brown bear categories, with the exception of areas with a rougher terrain that were used by all bears in IHY and EM (Fig. 2). Compared to males, females generally showed more homogeneous habitat use patterns throughout the year and used areas closer to human settlements in eHY and lHY; adult males showed the highest seasonal variation in their proximity to most habitat features (Fig. 2). Adult males used shrublands and the most natural areas, such as higher-elevation areas and areas closer to grasslands from eHY to EM and areas closer to water courses in PM. From MA to eHY, adult and subadult males used areas closer to anthropogenic habitat features, such as human settlements and naturalized and intensive crops (Fig. 2).

In general, all bear categories used areas closer to natural habitat features (e.g. forests, shrublands, areas with high altitude and rough terrain) during the day and areas closer to human-related habitat features (e.g. intensive and naturalized crops, human

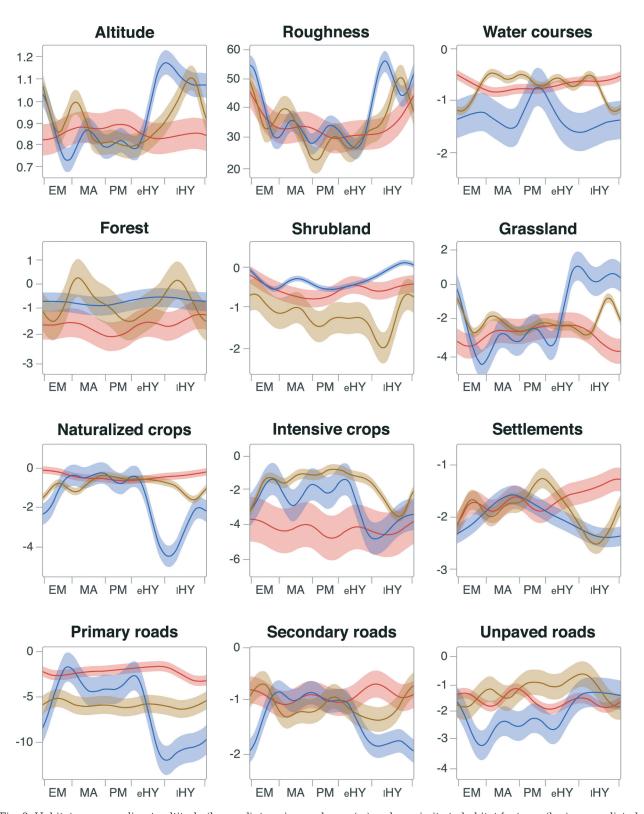


Fig. 2. Habitat use according to altitude (km a.s.l), terrain roughness (m) and proximity to habitat features (km), as predicted by GAMMs (y-axis) for adult female (red line), adult male (blue line) and subadult (brown line) brown bears in Greece across the biological seasons (x-axis) (EM: emergence; MA: mating; PM: post-mating; eHY: early hyperphagia; lHY: late hyperphagia). Shaded areas indicate the standard deviation for each bear category. The smooth terms for all models are significant at p < 0.001

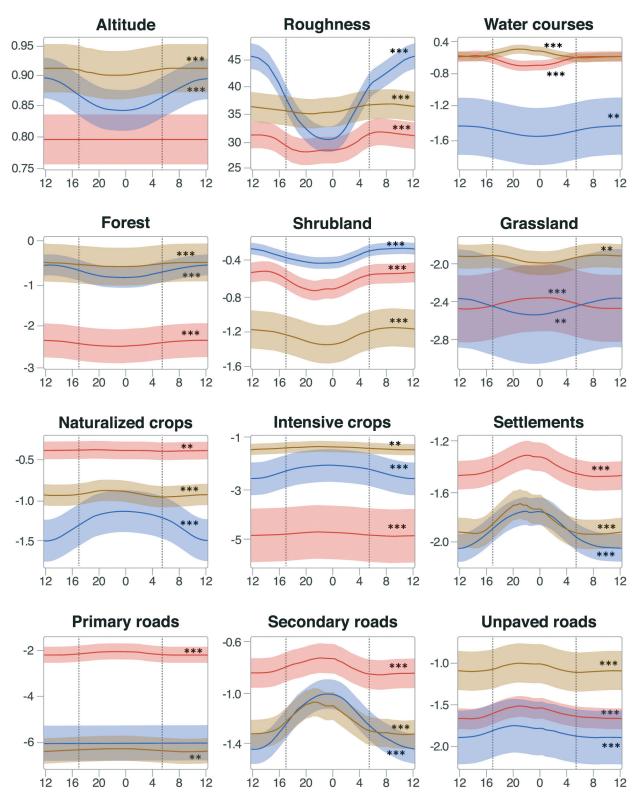


Fig. 3. Habitat use according to altitude (km a.s.l), terrain roughness (m), and proximity to habitat features (km), as predicted by GAMMs (y-axis) for adult female (red line), adult male (blue line) and subadult male (brown line) brown bears in Greece across hours of the day (x-axis). Shaded areas indicate the standard deviation for each age-sex class. Dotted vertical lines indicate the average sunset and sunrise from the emergence season through to the late hyperphagia season. The significance levels of the smooth terms are \*\*\*p < 0.001; \*\*p < 0.01; and \*p < 0.05

settlements, primary, secondary and unpaved roads) during the night (Fig. 3). Water courses were approached during the day by adult females and males and during the night by subadult males; areas close to grasslands were used during the day by subadult and adult males and during the night by females. Overall, adult males had a more pronounced daynight shift in their proximity to most of the habitat features (Fig. 3).

# 3.2. Habitat selection for stationary and moving behaviors

In general, bears showed a positive selection for areas with rough terrain and naturalized crops and areas close to water courses while avoiding high-altitude areas and primary and unpaved roads (Fig. 4). However, habitat selection showed distinct differences across bear categories and behavioral statuses; adult males selected for forests, shrublands and grasslands, which were avoided or not selected by females and subadult males. Adult males also selected for areas close to human settlements, which were avoided by adult females. Intensive crops were selected by adult females and subadult males for moving behaviors and were avoided by adult and subadult males for stationary behaviors. Secondary roads were selected by adult females for moving behaviors, but avoided by males for all behaviors (Fig. 4; see Tables S2 & S3 for details on model selection and GLMM estimates).

# 4. DISCUSSION

Understanding habitat use and selection patterns is of paramount importance in wildlife conservation, as behavioral responses may often be the first measurable reactions that animals show in response to (human-induced) environmental changes and can help determine the capacity of a species to adapt to these changes (Sih et al. 2011). In contrast to most studies that have been carried out in the northern and central parts of Europe, our study provides new contributions to understanding the habitat ecology of brown bears by examining them in the anthropogenic landscape of southern Europe.

Overall, the results of the study indicate that habitat use and selection by bears in Greece favored habitats with low anthropogenic disturbance that are suitable for covering the biological requirements of the species (i.e. including nutrition, refuge, hiberna-

tion and reproduction) (Naves et al. 2003, Posillico et al. 2004).

All bears in this study used areas with high topographic complexity (i.e. areas with rough terrain) in lHY and EM. The use of such areas at this time of the year is a behavior that has been recorded throughout the range of the species (e.g. Europe: Swenson et al. 2020; North America: Apps et al. 2004) and has been associated with better protection from humans, better sheltering and denning opportunities (Linnell et al. 2000, Nielsen et al. 2004) and the existence of complementary feeding habitats (May et al. 2008, Güthlin et al. 2011). The increased variability in circannual habitat use by males was likely a result of their higher mobility (de Gabriel Hernando et al. 2020) that enabled them to adapt more dynamically to the spatio-temporal changes in resource and refuge availability of the fragmented landscape in Greece. Furthermore, males used areas closer to anthropogenic habitat features from MA to eHY than during the rest of the year, likely in an attempt to make use of the primary food sources that are available at this time of the year: the diet of bears in Greece in the summer includes a relatively high proportion of cereals and soft mast (Paralikidis et al. 2010), which are found mainly in anthropogenic landscapes. This behavior is in accordance with previous information (Mertzanis et al. 2008) and is consistent with the patterns of human-bear conflicts (i.e. damage to crops) in the country (Karamanlidis et al. 2011).

Brown bear habitat use in Greece had a clear circadian pattern, i.e. closer to natural habitats during the day and closer to human habitats during the night. This pattern was similar across all bear categories, with the exception of proximity to grasslands and water courses. Circadian habitat use patterns of brown bears have been recorded previously in North America (Nielsen et al. 2004) and in Scandinavia, where human activity did not appear to have relevant effects (Ordiz et al. 2014), probably because human densities in Scandinavia are among the lowest within bear ranges in Europe (Ordiz et al. 2011). In contrast, studies in areas with higher human density, such as Greece (Mertzanis et al. 2008) or Spain (Naves et al. 2001), have indicated that human activity may have a significant effect on circadian habitat use. Consequently, the use of natural habitats by females and subadult males during the day should be regarded as a compromise between avoiding human activity and intraspecific aggression or competition with adult males while satisfying higher nutritional needs and lack of experience, respectively (Elfström

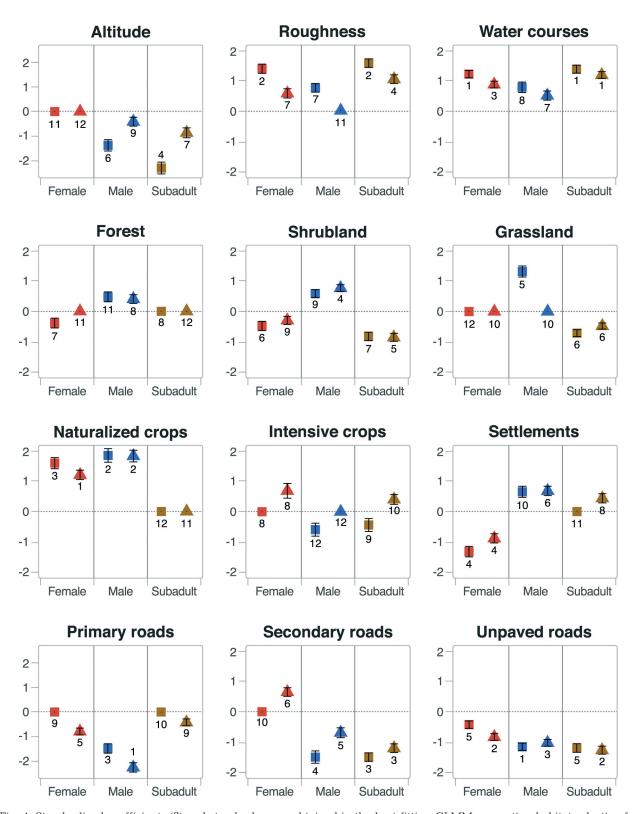


Fig. 4. Standardized coefficients ( $\beta$ ) and standard errors obtained in the best-fitting GLMM supporting habitat selection for stationary behaviors (squares) and moving behaviors (triangles) for adult female (red), adult male (blue) and subadult male (brown) brown bears in Greece. Symbols at the y-axis value of 0 represent variables that did not enter the best-fitting models for each behavior and bear category. Variable importance ranking according to  $\Delta$ AIC, where 1 is the most important, is represented by the numbers next to each feature

et al. 2014a,b). In regard to the aforementioned exceptions, the daytime use of areas closer to water courses by adult bears in Greece suggests an attempt to cope with the high ambient temperatures of the Mediterranean landscape (de Gabriel Hernando et al. 2020). In addition, females were closer to grasslands during the night, while males were closer to this habitat feature during the day, which could be the result of efforts by females to avoid males (Steyaert et al. 2016) while profiting from their use as important feeding areas (Mealey 1980).

Brown bears in Greece showed consistent habitat selection patterns across all bear categories for specific habitat features: they selected for areas with rough terrain and close to water courses, while avoiding primary and unpaved roads. This behavior should be interpreted as an attempt to avoid human disturbance (Nellemann et al. 2007), which is common among bears (Naves et al. 2003, Zarzo-Arias et al. 2018). Selection towards naturalized crops was also generally consistent across adult bears, highlighting their importance as feeding areas and/or movement corridors with low or moderate human disturbance (Ambarlı 2012). We believe that the increase in the availability of such areas following the rural abandonment in Greece in combination with the use of shrublands by adult males has favored the dispersal of individuals in the country (Karamanlidis et al. 2021) and has been one of the main drivers of the spatial expansion and recovery of the brown bear in Greece (Bonnet Lebrun et al. 2020).

Differences in habitat selection for specific habitat features across bear categories might be due to the ability of females and subadults to anticipate habitat selection by adult males (Steyaert et al. 2016) and, consequently, avoiding them. For example, forests, shrublands and grasslands, which combined are considered as optimal bear habitat (Nielsen et al. 2004), were avoided by females and subadults. Similarly, adult females keeping away from human settlements might be viewed as an attempt to avoid dominant males that use these areas in search of human-associated food resources; this behavior of adult males could be the result of a progressive loss of fear in the context of rural abandonment and decreased human persecution (Martínez-Abraín et al. 2019).

Habitat selection by brown bears in Greece varied between stationary and moving behaviors, which is not uncommon for bears in Europe (e.g. Scandinavia; Moe et al. 2007). These variations provide insights into the ecological role of the various habitat features: higher selection for a specific habitat for stationary behaviors (e.g. areas with rough terrain or

water courses) could be considered as an indication of this specific habitat serving as refuge habitat (Moe et al. 2007), while higher selection for moving behaviors (e.g. shrublands) might indicate suitability as a movement corridor (Dickie et al. 2020). However, some disturbed habitats, such as areas close to human settlements that were selected by male bears or secondary roads that were selected by females, could act as attractive sinks due to a poorly perceived risk of human disturbance (Naves et al. 2003, Penteriani et al. 2018).

The results of the study provide detailed information on the ecological role of various habitat types that we recommend should be used to develop guidelines for species conservation and allow for prioritizing management actions that will promote the conservation of bears in Greece. At the same time, this information may also contribute to promoting the recovery/persistence of other (endangered) species inhabiting anthropogenic landscapes.

The circannual and circadian habitat use patterns provide information on how to limit interactions between bears and humans in space and/or time: it is of utmost importance to restrict human activity in natural areas which act as refuge areas (e.g. forests in areas with rough terrain) during the day in IHY and EM. In addition, the habitat selection patterns indicate the importance of (1) reducing human activity in bear habitat and (2) protecting/improving the condition (i.e. quantity and/or quality) of specific natural habitat types in order to protect bears (Pop et al. 2018). With regard to reducing human activity in bear habitats, this may be most important for suitable habitat positively selected in spring and summer, as during this period, bears are closer to human features and appear to use smaller areas (Berland et al. 2008), and because this is a critical time for cub survival (Planella et al. 2019).

Our results indicate that within the context of the anthropogenic Mediterranean landscape of Greece, specific types of habitat are critical for the conservation of the species. These include shrublands, grasslands and naturalized crops because of their importance as potential feeding and/or corridor areas, while areas with rough terrain and water courses appear to be highly important as refuge areas. The latter habitat feature is particularly important, especially when considering that it is already heavily impacted in the country (Dimopoulos et al. 2006) and that the effects of climate change in Greece are expected to be high (Giannakopoulos et al. 2009). On the other hand, avoidance of certain types of habitat should also inform bear conservation. In this respect,

roads appear to be particularly important; the importance of roadless areas for the conservation of wildlife is globally acknowledged (Ibisch et al. 2016, Penteriani et al. 2018, Morales-González et al. 2020). From a management perspective, particular attention will also need to be given to high-risk areas (i.e. areas with increased human activity that appear to be attractive to bears such as the surroundings of human settlements, secondary roads or intensive crops) in order to prevent them from functioning as ecological traps (Northrup et al. 2012, Penteriani et al. 2018, Morales-González et al. 2020). Finally, safequarding natural areas of suitable habitat that can act as corridors of dispersal and gene flow will be important for successful recolonization (Nellemann et al. 2007), the establishment of metapopulations and the reconnection of fragmented population units (Karamanlidis et al. 2018).

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